

Assessment of Heavy Metals in Road Dust of Behbahan City, Iran: Distribution, Sources and Health Risks

Zeinab Ghaedrahmat^{1,2}, Halimeh Almasi^{1,2}, Razegheh Akhbarizadeh³, Mehdi Ahmadi^{4*}

¹ Department of Environmental Health Engineering, Shoushtar Faculty of Medical Sciences, Shoushtar, Iran.

² Student Researcher Committee, Shoushtar Faculty of Medical Sciences, Shoushtar, Iran.

³ System Environmental Health and Energy Research Center, The Persian Gulf Biomedical Sciences Research Institute, Bushehr University of Medical Sciences, Bushehr, Iran.

⁴ Environmental Technologies Research Center, Ahvaz Jundishapur University of Medical Sciences, Ahvaz, Iran.

ARTICLE INFO

ORIGINAL ARTICLE

Article History:

Received: 19 March 2022

Accepted: 20 May 2022

*Corresponding Author:

Mehdi Ahmadi

Email:

Ahmadi241@gmail.com

Tel:

+989126779273

Keywords:

Dust,
Metals, Heavy,
Risk Assessment,
Risk Factors,
Behbahan City.

ABSTRACT

Introduction: Road dust is a group of solid particles that are presented in the urban areas and are originated from both natural and human-induced sources. This study aims to determine concentrations of heavy metals in urban dusts in Behbahan to identify their natural or anthropogenic sources.

Materials and Methods: In this study, a total of 20 samples were collected from main roads with different land uses, including residential, industrial, and commercial areas in Behbahan city, Khuzestan province.

Results: The results of mineralogy identification showed that calcite, dolomite, quartz, albite, and gypsum were the most identified minerals in dust samples. In addition, heavy metals of As, Co, Cr, Cu, Mn, Fe, Ni, Zn, Pb, Sb, U, Cd, Hg, and Mo were investigated in the collected settled road dust. The results indicated that the level of heavy metals, such as Co and Sb in the places with high traffic load were much higher than other areas. The highest I_{geo} value for Pb, Cu, and Sb were 1.39, 2.19, 2.46, respectively, suggesting that Pb, Cu, and Sb in the road dust were originated from the anthropogenic sources. Moreover, the results demonstrated that road dust may pose serious health threats to humans (both adults and children).

Conclusion: In this study, the concentration of heavy metals in road dust of Behbahan was investigated. The concentration of heavy metals, such as Cu, Zn, Ni, As, Cr, Pb, U, and Fe in the road dust of the commercial section were much higher than other sections.

Citation: Ghaedrahmat Z, Almasi H, Akhbarizadeh R, et al. *Assessment of Heavy Metals in Road Dust of Behbahan City, Iran: Distribution, Sources and Health Risks*. J Environ Health Sustain Dev. 2022; 7(2): 1632-46.

Introduction

The road dust is a group of solid particles that accumulate in urban areas from both natural and human-induced sources. It should be noted that human resources have a greater share in creating road dust. Road dust particles can be considered as a main source of heavy metals¹⁻⁴. Natural and human-induced airborne particles have a great effect on human health and environment. These particles (dust) are very common phenomena in

sizes ranging from 1 to 10,000 μm . Dust generates from a variety of sources, such as mining, is settled due to their high resistance, toxicity, and accumulation in organisms⁵.

The origin of heavy metals in dust particles of the urban areas are mostly traffics and possible industrial activity in the vicinity of city. In addition, heavy metals in the dust particles are chemically composed of silicates, carbonates, and organic matter. The accumulated pollutants

in the road dust have a short shelf life; so their amounts in the settled road dust reflect the recent contamination. Therefore, road dust should be considered as an important indicator for measuring recently accumulated pollutants (i.e. PTEs) ⁶⁻⁹.

The main exposure pathways for dust and their associated chemicals, such as heavy metals are inhalation, ingestion, and skin contact. The interned metals may accumulate in the body's adipose tissue, build on the central nervous system, accumulate in different organs of the human body, and disrupt the function of these organs. Therefore, detection and estimation of exposure to heavy metals through road dust is very important ^{4, 10}. The smaller dust particles are greater risk. Particles smaller than 100µm are easily suspended and enter the breathing system, and particles smaller than 10µm can enter the lungs. Inhalable dust smaller than 2.5µm can enter the blood stream. Particles with a diameter of 1 to 10µm, especially particles with a diameter of less than 2.5µm, have the most harmful effects on human health. It should be noted that too small particles do not have enough time to suck in the lungs and exhale after exiting the lung. The vehicles emissions in the urban areas involve very small particles that can be deposited on road surfaces or on impermeable surfaces. On the other hand, dust that enters the surrounding environment in urban area is also a factor in the transmission of pollutants. When

precipitation occurs, these particles are washed with various pollutants, and finally they are transferred to water in contact with them, such as urban runoffs ¹¹⁻¹³, which in all of these cases, creates health problems. This study aims to determine mean concentrations of heavy metals in urban dusts in Behbahan city to identify their natural or anthropogenic sources.

Materials and Methods

Study area

This study was conducted in the urban area of Behbahan city in south western of Khuzestan province in Iran with total area of 3195 km² (Figure 1). The population of Behbahan was 186293 in 2019. Its coordinates are 30.5959°N – 50.2417°W. Behbahan climate is semi-desert or foothills. Behbahan city due to its location in the proximity to Zagros Mountains and southern ports of Iran, as well as having rugged rivers of Maroon, Kheirabad, and Zohreh has special status and special areas. Pazanan, Mansourabad, and Peranj oilfields are located 2 km from Behbahan city. Also Bidband Gas refinery is located in 2 kilometers west of Behbahan city. Nowadays, due to different oil and gas industries around the city and other small and large industries in the city center, compliance with environmental laws and standards by industry owners is inevitable, in which regard the role of city environmental monitoring .It is very important and effective in reducing the amount of environmental pollution.

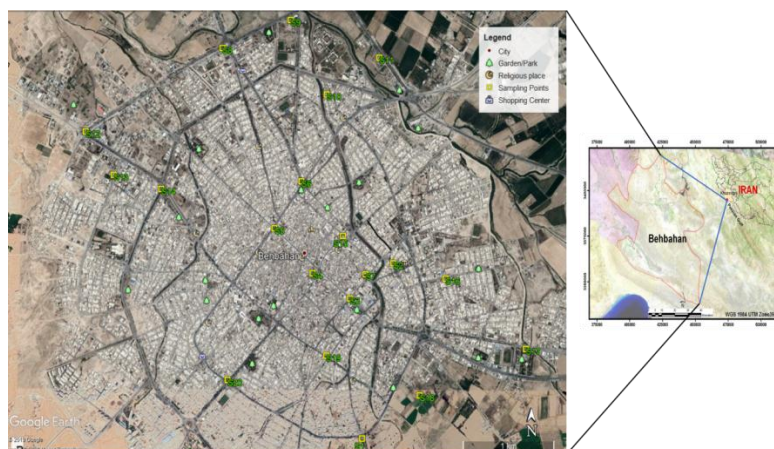


Figure 1: Location of sampling in Iran and sampling points of road dust in Behbahan city

The study area was divided into various land uses, including residential, industrial, commercial, and heavy traffic, and samples were taken from each user. The samples (n = 20) (Figure 1) were collected from main roads and main fields (due to road length and traffic volume), residential, industrial and commercial areas. In order to obtain more accurate data for risk assessment, the samples were collected from crowded places, such as hospitals, schools, parks, shopping centers.

Since road dust sampling was not feasible during the day, it was done after midnight. The samples were taken off the road, the road and the fields using a plastic sander, and plastic brush (between 300 and 500 grams of samples). For sampling with less error, the samples were taken as composite, so that each sample represented about 10 sub-samples. After each sampling, the brush was completely cleaned and washed with distilled water and acetone. The samples were then transferred to zipper thick bags and stored in a cool and dry environment after encoding. First, for separation of wooden and metal parts, leaves, cigarette filters, glass, and other unwanted materials from the samples, samples were taken from a 2 mm sieve, and then passed through a 63-micron beaker. The samples were then transferred to small, colorless, thick bags, encoded, and sent to laboratory for further analysis.

Sample preparation

The samples were mixed, dried, and passed through a 0.125mm sieve. The samples were digested with HNO₃, HF, and HClO₄ and concentrations of metals were analyzed with ICP/MSS, which its limit of quantification (LOQ) was in the range of 0.1-1000 ppm. The composition of road dust was determined with X-ray diffraction (Philips-expert-pro).

Statistical analysis

The EPA PMF 5.0.14, XLSTAT software (2016), and Microsoft Excel 2016 were applied. Shapiro-Wilk test showed that the data were not normal. Spearman correlation analysis was applied to evaluate the association between heavy metals.

Biplot principal component analysis (PCA) was used to clarify the relationship and sources of heavy metals. The significance level of all analyses was considered < 0.05.

Heavy metal evaluate in road dust

Contamination factor (CF), degree of contamination (DC), and pollution load index (PLI) were applied to evaluate heavy metal concentration in road dust, as well as to provide a criterion for determining the DC¹⁴. The CF, DC, and PLI parameters were calculated using Eqs.1, 2, and 3:

$$CF = \frac{C_{metal}}{C_{background}} \quad (1)$$

$$DC = \sum CF \quad (2)$$

$$PLI = \sqrt[n]{(CF_1 \times CF_1 \times CF_1 \times \dots \times CF_n)} \quad (3)$$

Where, C_{metal} is pollutant concentration in the dust, C_{Background} is amount of background metals, and n is the number of metals. The PLI is a comparison tool for evaluating the quality of site under investigation. The PLI index was presented by Tomilsson et al.¹⁵.

Pollution indices

The ecological risk index (ERI)¹⁴ and geo-accumulation (I-geo)¹⁶ are applied to assess the ecological risk of heavy metals in road dust. Geo-accumulation index (I_{geo}) was applied to specify the severity of heavy metal contamination of soil. This index was calculated using Eqs.4,5, and 6:

$$C_f = \frac{C_s}{C_n} \quad (4)$$

$$ERF = T_r \times C_f \quad (5) \quad RI = \sum_{i=1}^m ERF \quad (6)$$

Where, C_f is metal contamination index, C_n refers to metal concentration in the sample, C_s represents background metal concentration, ERF indicates potential ecological risk factor (ERF). T_r shows heavy metal response coefficient determined by Hackenson for different elements (Zn = 1 < Cr = 2 < Cu = Ni = Pb = 5 < As = 10 < Cd = 30 < Hg = 40), and ERI is ERI. I-geo was determined according to Eq.7:

$$I\text{-geo} = \log_2 \left(\frac{C_s}{1.5 \times B_n} \right) \quad (7)$$

Where, Cs is metal concentration in soil samples and Bn is geochemical background. 1.5 shows a correction factor for lithospheric effects ¹⁶.

The risk of total non-carcinogenic metals (As, Co, Cr, Cu, Mn, Ni, Zn, Pb, Sb, and U) and total carcinogenic risk for As, Co, Cr, Ni, and Pb can be determined by adding the risks of calculated exposure. The exposure dose for all three exposure routes (swallowing, breathing, and skin contact of dust particles) is calculated using Eq.8, 9, and 10 ¹⁷:

$$D_{ing}(mg\ kg^{-1}day^{-1}) = C(mg\ kg^{-1}) \times \frac{R_{ing} \times EF \times ED}{BW \times AT} \times 10^{-6} \quad (8)$$

$$D_{inh}(mg\ kg^{-1}day^{-1}) = C(mg\ kg^{-1}) \times \frac{R_{inh} \times EF \times ED}{PEF \times BW \times AT} \quad (9)$$

$$D_{dermal}(mg\ kg^{-1}day^{-1}) = C(mg\ kg^{-1}) \times \frac{SA \times SL \times ABS \times EF \times ED}{BW \times AT} \times 10^{-6} \quad (10)$$

The lifetime average daily dose (LADD) is used to assess cancer risk, which is found for respiratory elements (Cd, Co, Cr and Ni). It was calculated using Eq.11 ¹⁷.

$$LADD(mg\ kg^{-1}day^{-1}) = \frac{C(mg\ kg^{-1}) \times EF}{AT \times PEF} \times \left(\frac{CR_{child} \times ED_{child}}{BW_{child}} \times \frac{CR_{adult} \times ED_{adult}}{BW_{adult}} \right) \quad (11)$$

Where, R_{ing} is ingestion rate ¹⁷. The R_{inh} is the rate of respiration that accounts for $7.6\ m^3 day^{-1}$ for children and $20\ m^3 day^{-1}$ for adults. The EF is repeatedly exposed and is given 180 days of a year ^{6,18}. The ED is subject to year-round exposure, with values set by the US Environmental Protection Agency (EPA) being 6 years for children and 24 years for adults ¹⁷. The SA is an area of exposed skin that is $2800\ cm^2$ for children and $5700\ cm^2$ for adults. The SL is a skin aeration factor that is 0.2 for children and 0.07 $mg\ cm^{-3}$ for adults ¹⁷.

The CR is the rate of contact or absorption, CR for ingestion is Ing R, for respiration is Inh R, and for skin absorption is according to Eq.12:

$$CR = SL \cdot ABS \cdot SA \quad (12)$$

The ABS is the skin absorption coefficient for all elements except 0.001. This factor for arsenic is 0.03. The VF is the evaporation coefficient ¹⁷. The BW is the average body weight (Kg) that is 15 for children and 70 for adults ¹⁷. The AT is the average time that is 25550 days for carcinogenic elements and $ED \times 365$ for non-carcinogenic elements ⁶. The C parameter is also the concentration of exposed elements.

Based on this parameter the hazard quotient (HQ) as well as the hazard index (HI) can be calculated ¹⁷. The HQ is based on the semantic risk of non-carcinogenic elements, and is calculated by dividing the average daily absorption by reference dose (RfD) (Man et al., 2010). The RfD ($mg\ kg^{-1} day^{-1}$) is an evaluation of the maximum allowable daily intake in human life. The HI is the sum of HQs and is used to evaluate health risk for exposure from various routes (Eqs.12, 13, 14, and 15). If the HI value ≤ 1 , it will not adversely affect human health and if the HI value > 1 , it will be detrimental to human health ¹⁹.

$$HQ_{ing} = \frac{D_{ing}}{oral\ RfD} \quad (12)$$

$$HQ_{inh} = \frac{D_{inh}}{inhal\ RfD} \quad (13)$$

$$HQ_{dermal} = \frac{D_{dermal}}{dermal\ RfD} \quad (14)$$

$$HI = \sum HQ \quad (15)$$

For carcinogens, carcinogenic risk (CR) is also estimated by multiplying the dose in the carcinogenic process ⁶.

$$Cancer\ Risk = LADD \cdot SF \quad (SF = Slope\ Factor) \quad (16)$$

Ethical issue

The ethical issue of this research was IR.NIMAD.REC.1397.009.

Results

XRD analysis

The Minerals, such as calcite, dolomite, quartz, albite, lizardite, lithite, gypsum, rutile, arsenolite, cristobalite, and muscovite were identified in the road dust samples of the study area (Figure 2). Calcite, dolomite, and quartz minerals were the dominant minerals in the sampling point. All minerals were distributed equally in all sampling stations.

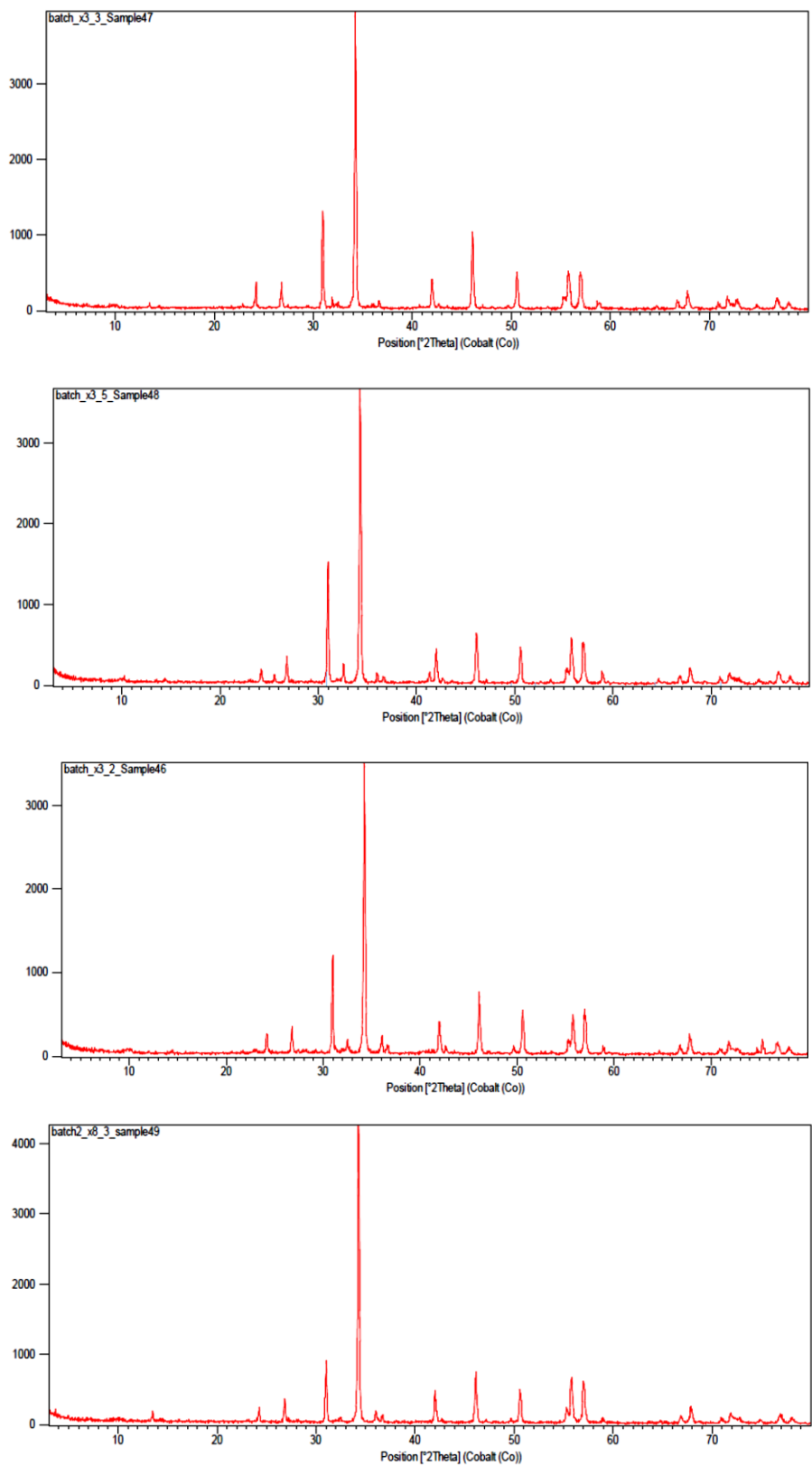


Figure 2: XRD analysis of road dust

Heavy metal concentration in road dust

Road dust samples for As, Co, Cr, Cu, Mn, Fe, Ni, Zn, Pb, Sb, U, Cd, Hg, and Mo were investigated to evaluate metal dust contamination from different sampling sites. Among these metals, Cd and Hg were excluded from the statistical analysis, since their concentrations in all samples were below the detection limit in the dust samples. The descriptive statistics of heavy metals concentrations in road dust of Behbahan city were showed in Table 1. The coefficient of variation of the concentration of elements for Cu and Pb is lower than 50%, indicating the low variation of the concentration of these elements in the soil of the region. The sampling points can be subdivided into 4 different zones, including residential, commercial, industrial, and heavy traffic. The heavy metals level in each section

were as follows: residential zone $Mn > Zn > Pb > Cu > Cr > Ni > As > Co > Mo > Sb > U > Fe$, commercial zone $Mn > Zn > Pb > Cu > Cr > Ni > As > Co > Mo > Sb > U > Fe$, industrial zone $Mn > Pb > Zn > Cr > Cu > Ni > As > Co > Sb > Mo > U > Fe$, and heavy traffic $Mn > Zn > Cu > Pb > Cr > Ni > As > Sb > Co > Mo > U > Fe$.

The concentration of heavy metals, such as Cu, Zn, Ni, As, Cr, Pb, U, and Fe in the road dust of commercial zone were much higher than other zones due to the most heavy traffic in this zone. The concentration of heavy metal, such as Co and Sb in the heavy traffic zone were much higher than other zones. The I_{geo} value was presented in Figure 3, the main I_{geo} value was showed I_{geo} value lower than 0, indicating that road dust in Behbahan city was uncontaminated with Co, Cr, Fe, Zn, U, Mo, Ni, and Mn.

Table 1: Descriptive statistics of elements concentration (mg/kg) of road dusts in Behbahan city

Statistic	As	Co	Cr	Cu	Fe	Mn	Mo	Ni	Pb	Sb	U	Zn
No. of samples	20	20	20	20	20	20	20	20	20	20	20	20
Minimum	5.44	2.43	21.61	14.00	0.57	104.39	1.21	8.99	12.44	1.13	0.75	32.95
Maximum	13.97	4.54	58.23	95.69	1.2	169.35	3.71	18.39	98.38	5.11	2.07	126.19
1st quartile	6.89	3.18	26.26	20.50	0.7675	118.04	1.31	11.70	22.04	2.06	1.04	52.20
Median	7.36	3.50	27.89	26.74	0.82	129.56	1.55	13.00	36.89	2.36	1.26	66.32
3rd quartile	8.57	3.67	31.77	44.36	0.945	132.96	2.22	14.49	56.83	2.87	1.47	82.44
Mean	8.04	3.50	30.08	35.70	0.8525	129.63	1.79	13.12	42.16	2.50	1.25	68.94
Standard deviation	1.86	0.55	7.45	22.18	0.1575	18.36	0.63	2.25	25.62	0.78	0.36	24.15
CV(SD/median)	0.23	0.16	0.25	0.62	0.18	0.14	0.35	0.17	0.61	0.31	0.29	0.35
Skewness	1.57	0.12	2.59	1.32	0.36	0.81	1.47	0.30	0.81	1.59	0.17	0.60
Background value	4.7	6.9	42	14	3.5	412.4	1.8	18	25.02	0.62	3.7	62

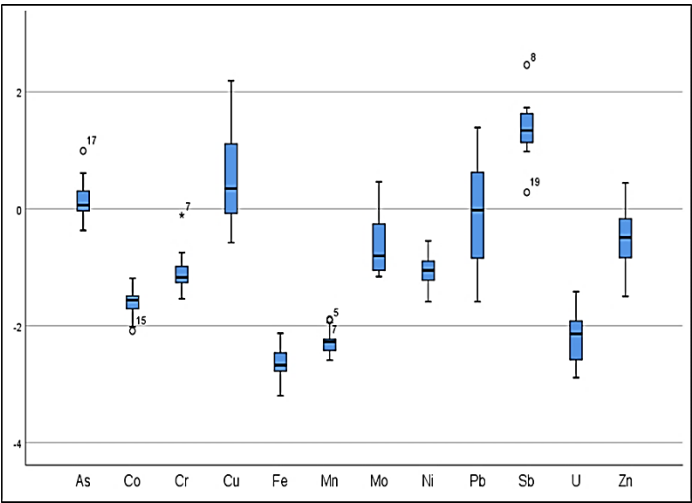


Figure 3: I_{geo} of road dusts in Behbahan city

Geo-accumulation index

The I_{geo} index value for Sb was 1.3, which showed that road dust in Behbahan city was moderately contaminated with Sb. The highest I_{geo} value for each metal also indicated that the road dust was partially contaminated and uncontaminated to moderately contaminated by the metals except for Pb, Cu, and Sb.

Contamination load index

The values of CF (Figure 4), DC, and PLI were calculated. The results showed that CF values of Co, Cr, Fe, Mn, Mo, and U were in the scope of low CF ($CF < 1$), while CF values of Ni, Cu, Pb, and Zn were in the scope of moderate contamination ($1 < CF < 3$). The CF value of Sb was in the scope of considerable contamination ($3 < CF < 6$).

In the evaluation of the CF index, Sb, Pb, and Cu had the most pollution in all stations, respectively. The DC value in all stations was in the scope of moderate contamination ($8 \leq DC \leq 16$) except stations 3, 7, 8, and 12. Based on the results, most stations were classified as uncontaminated sites ($PLI < 1$) and stations 7 and 8 were classified as contaminated sites ($PLI > 1$). These two stations are located in heavy traffic areas of Behbahan city.

Ecological risk assessment

ERF for all metals was determined low with ERF values < 40 except Pb that ERF was determined moderate (Figure 5). ERI in all sectors was determined low with ERI values < 150 .

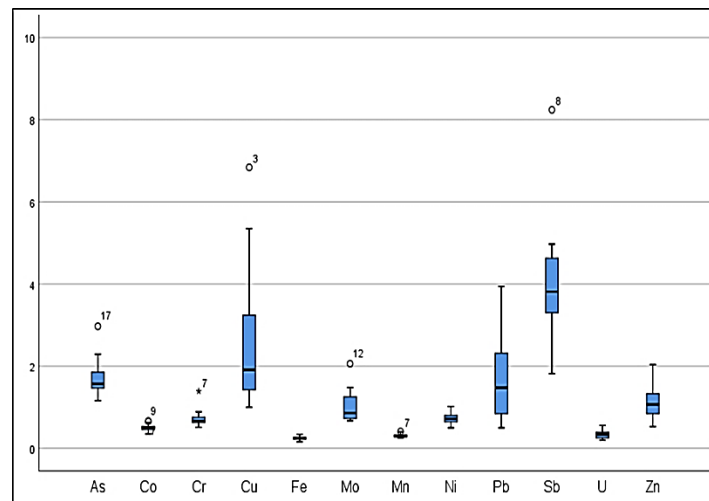


Figure 4: The CF of road dusts in Behbahan

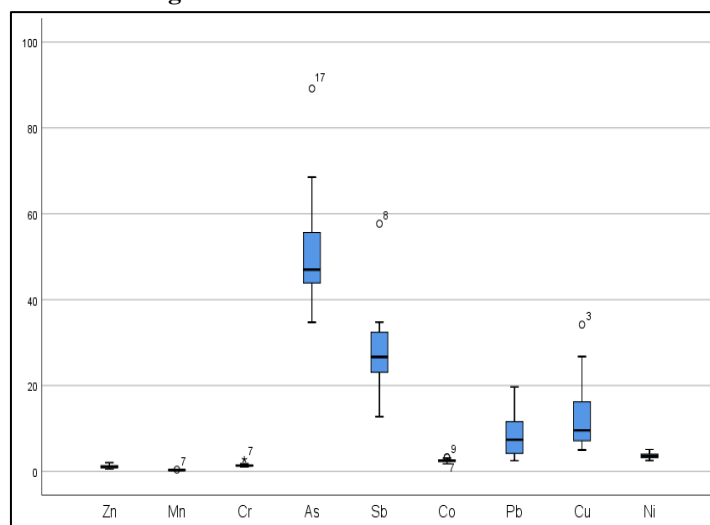


Figure 5: Determination of ERF in road dusts of Behbahan

The results of spearman correlation coefficients are presented in Table 2 for metals in road dust samples of Behbahan city. Specific positive correlations were considered for some metals: As-Sb, As-U, Co-Cr, Co-Cu, Co-Fe, Co-Mn, Co-Ni, Co-Zn, Cr-Cu, Cu-Fe, Cu-Mn, Cu-Ni, Cu-Zn, Fe-Mn, Fe-Zn, Mn-Ni, Mn-Zn, Mo-Sb, Ni-Zn, Pb-Sb.

In addition, the PC Biplot demonstrates the spatial distribution of the sampling stations according to their symmetry. The stations in the

Positive matrix factorization (PMF) model was performed to investigate the contribution of various sources of heavy metal in road dust of Behbahan city. Results were indicated that urban source (factor 1) was prevailing by Pb (35%), Cu (72.5%) and Zn (54.6%) (Figure 7).

Table 2: Spearman correlation analysis for metals concentrations of road dust in Behbahan

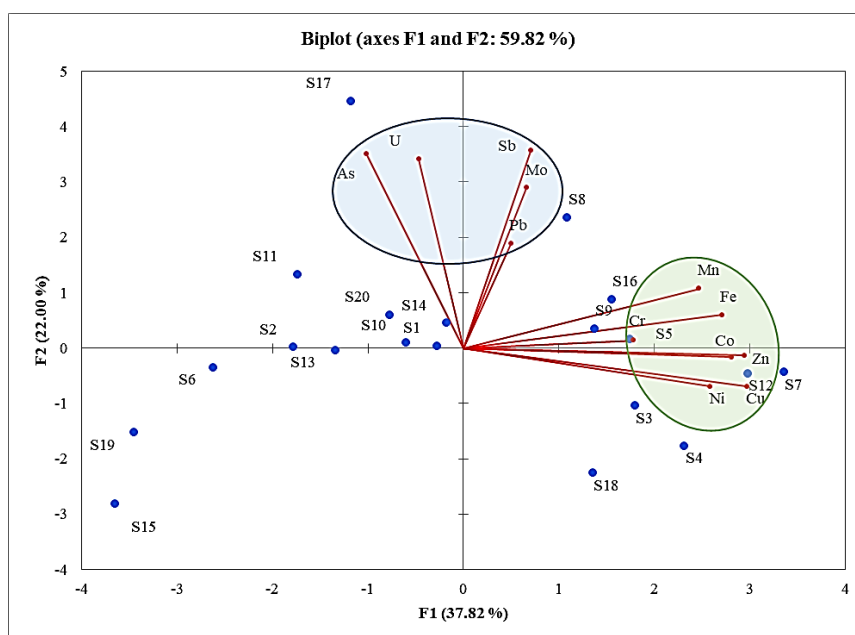
[illegible]

Figure 6: PC biplot of station and metals in road dust of Behbahan

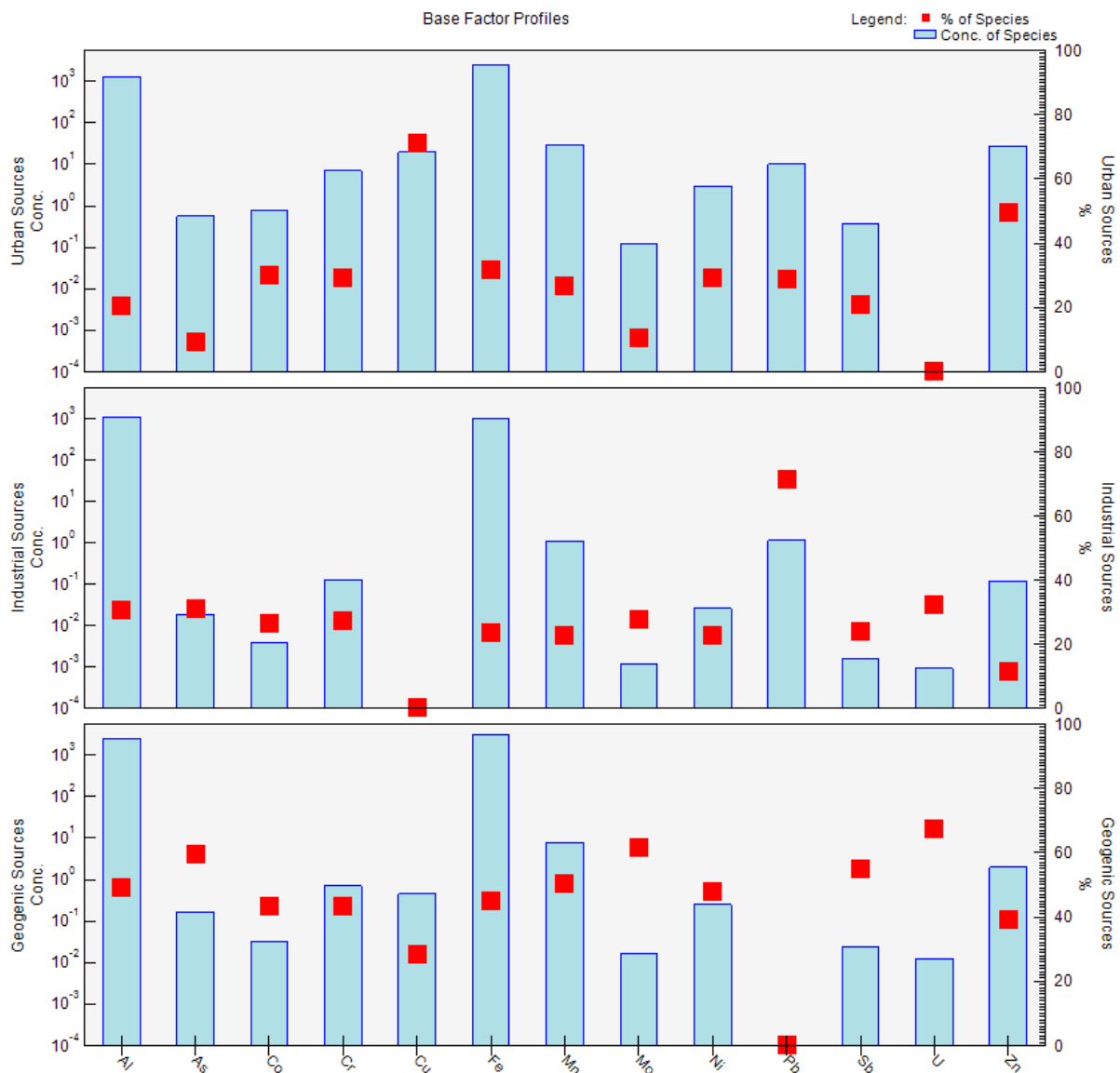


Figure 7: Source profile and factor fingerprint of the road dust samples from PMF model analysis

Health risk assessment

The health risks for humans caused due to heavy metals in road dust of Behbahan city, through various pathways (ingestion, inhalation, and dermal contact), were calculated (Table 3). The results indicated that chronic exposure of humans (both adults and children) to road dust in the study area may pose adverse health effects (HQ, HI, and THI > 1).

However, the sequence of non-carcinogenic risk of different exposure for children was dermal contact > inhalation > ingestion and for

adults were dermal contact > ingestion > inhalation. Moreover, lower HQ values were obtained for adults in all metals and through all exposure pathways. Hence, children may have more potential non-carcinogenic risks than adults. In the case of carcinogenic risks, as shown in Table 3, the average carcinogenic risks of carcinogen metals (As, Co, Cr, Ni, and Pb) under different routes for adults and children were below 10^{-2} . Furthermore, TCR for adults and children were $6.89\text{E-}03$ and $4.02\text{E-}02$, respectively.

Table 3: Health risks due to environmental exposure to road dust's metals in Behbahan city

Element/index		Children			Adults		
		Inhalation	Ingestion	Dermal	Inhalation	Ingestion	Dermal
Non-carcinogenic risk							
HQ	As	-	4.3	3.46	-	8.56	7
	Co	-	2.8×10^{-1}	1.26×10^{-1}	6.6×10^{-2}	3.25×10^{-2}	1.8×10^{-2}
	Cr	1.78	1.3	1.66	4.6×10^{-2}	1.86	4.15
	Cu	-	1.425	1	-	1.65×10^{-1}	2.9×10^{-1}
	MO	-	5.72×10^{-1}	6.44×10^{-1}	-	6.6×10^{-2}	1.48×10^{-1}
	Ni	-	1.05	9.4×10^{-1}	-	1.22×10^{-1}	2.16×10^{-1}
	Pb	-	2.23	30	-	2.6	7
	Sb	-	10.05	11.12	-	1.16	25.87
	U	-	6.7×10^{-1}	-	-	3.3×10^{-2}	1.47
	Zn	-	3.6×10^{-1}	4×10^{-1}	-	4×10^{-2}	9.5×10^{-2}
HI	1.78	20.35	47.24	1.12×10^{-1}	14.18	45.49	
THI	69.375			59.67			
Carcinogenic risk							
CR	As	1.5×10^{-4}	1.9×10^{-2}	1.2×10^{-3}	2.3×10^{-8}	9.75×10^{-4}	1.061×10^{-3}
	Co	-	-	1.1×10^{-2}	-	-	2.84×10^{-3}
	Cr	-	-	4.2×10^{-3}	-	-	1.045×10^{-3}
	Ni	-	-	3.94×10^{-3}	-	-	9.07×10^{-4}
	Pb	2.1×10^{-4}	5.69×10^{-4}	-	7.77×10^{-8}	6.63×10^{-5}	-
TCR	4.02×10^{-2}			6.89×10^{-3}			

Discussion

The composition of road dust can vary depending on the factors, such as geographic place, land use, traffic specification, and drought periods^{3, 20}.

The differences in the amount of minerals at different stations can be due to their different environmental conditions. The percentage of calcite mineralization in downtown stations increased. This could be due to increased construction activities (using plaster, etc.) which are more in downtown than other areas. In areas of the city where there are more sources of pollution, various minerals (such as car exhaust) are released which can reduce the quartz percentage²¹. The iron oxides observed in road dust samples can be affected by anthropogenic sources in addition to terrestrial sources. Magnetite minerals, for example, are probably produced by the burning of fossil fuels, vehicle exhaust emissions, solid waste dispersal, and dust from smelting activities and other industries²².

The sampling points can be subdivided into 4 different sections, including residential, commercial, industrial, and heavy traffic. The heavy metals level in each section were as follows:

residential section Mn > Zn > Pb > Cu > Cr > Ni > As > Co > Mo > Sb > U > Fe, commercial section Mn > Zn > Pb > Cu > Cr > Ni > As > Co > Mo > Sb > U > Fe, industrial section Mn > Pb > Zn > Cr > Cu > Ni > As > Co > Sb > Mo > U > Fe, and heavy traffic Mn > Zn > Cu > Pb > Cr > Ni > As > Sb > Co > Mo > U > Fe. Similar trend was reported in previous studies²³. The mean concentrations of heavy metals, such as As, Cu, Zn, Pb, and Sb were higher than background heavy metal concentrations. The results were observed in road dust of Isfahan city²⁴.

A previous study reported that the mean concentrations of heavy metals were higher than background concentrations, indicating that the metals originate from human resources in the region²⁵. On the other hand, another study reported that it is not possible to conclude about soil pollution and the role of human activities on the concentration of metals only with respect to the background concentration. It is due the fact that the relationship between heavy metals in the soil is complex and ambiguous^{25, 26}. Many factors, such as the concentrations of heavy metals in rocks and parent materials, various soil formation processes, and human factors determine the

relative abundance of metal concentrations in soils²⁷.

The concentration of Mn and Mo in the residential section was much higher than other sections, suggesting the two metals in road dust mainly related to natural sources^{18, 28, 29}.

These metals vary extensively between different zones and between samples, due to heterogeneity of sample and different variables that cannot be controlled. This can be described by a high grade of metal variation regardless of distinctive characteristic of economic activity in each section^{23, 30}. Depending on coefficient of variation (CV), the studied metals can be divided into two categories, including metals with CV higher than 0.4, such as Cu and Pb and metals with CV less than 0.4, such as As, Co, Cr, Ni, Fe, Mn, Mo, Sb, and Zn. It is expected that metals with low CV are influenced by natural resources. However, metals with high CV are influenced by anthropogenic resources. This trend is similar to studies conducted in other cities of Iran, including Isfahan and Tehran^{24, 31}. The range of changes in the concentration of heavy metal and the CV of heavy metal concentrations indicate the human factors affecting the concentration of metals in the road dust of study area. The highest CV were related to Cu (0.62) and Pb (0.61), indicating the effect of human activities on increasing the concentration of these metals in the road dust of the region. The concentration of Mn, Cu, and Pb in all sections were significant. These metals are associated to road pavement materials, tire and brake wear, lubricants, and fuel combustion^{23, 28}. The Pb levels in areas with low traffic around industries section are significantly lower than other areas. It was shown that industry cannot play a specific role in the level of Pb in road dust and vehicle have a major role³².

The Zn level mainly originates from cars due to lubricating oils and tires³². Due to the high temperature, tire wear can be increased and Zn was also used as a welding agent in tires³³. The existence of Cu in road dust can be obtained from car engine wear. Also Brake pads are rich in metals, especially Cu³⁴. Fe is a widely used metal

in industrial practices and especially in smelting industries³⁵. Cr originates from the erosion of Cr coatings and alloys in automobiles. The main reasons for presence of Ni are the use of fossil fuels by industries in this region. The Ni and Mn mainly have industrial origin³⁶. The mean concentrations of heavy metals in Behbahan road dust were lower than other cities of Iran, such as Esfahan²⁴ and Tehran³¹.

The I_{geo} was applied to investigate the severity of heavy metal contamination of soil. The I_{geo} was introduced by Muller¹⁶ to assess soil contamination by heavy metals. This index is applied to assign the level of contamination and impact of human factors on natural factors in the soil and sediment environment. The I_{geo} index value was in the range of 0 to 1, indicating that road dust in Behbahan city was uncontaminated to moderately contaminated with As and Cu. As and its compounds are used as pesticides. Metal arsenic is also used in the production of alloys of Pb, Cu, and steel and in the electronics industry³⁷.

The I_{geo} value for Sb was 1.3, revealing that road dust in Behbahan city was moderately contaminated with Sb. This is due to its small amount of this metal in the earth's crust³⁸. The highest I_{geo} value for Pb, Cu, and Sb were 1.39, 2.19, and 2.46, which may suggest that Pb, Cu, and Sb in the road dust were most affected by anthropogenic sources. These results have been observed in other studies^{39, 40}. The Cu compounds are used as anti-corrosives in ores. These compounds create a protective layer on the surface of the engine that reduces friction and prevents damage from wear between different parts of the engine³⁷.

The high level of Pb in the road dust was usually associated with point sources of pollution and transportation of vehicle emissions. It was observed that increase in Pb concentration in road dust was caused by contamination of exhaust emission components and wear of car components²⁹.

The values of CF (Figure 4), DC, and PLI were calculated. In the evaluation of the CF index, Sb, Pb, and Cu had the most pollution in all stations,

respectively. The DC value in all stations was in the scope of moderate contamination ($8 \leq DC \leq 16$) except stations 3, 7, 8, and 12. This phenomenon is due to this fact that these areas are overcrowded and the pollution caused by vehicles in these areas is very high.

The lowest ecological risk was determined in residential zone. Heavy traffic zone shows the maximum ERI value and ecological risk of the road dust samples. This phenomenon is the main local source of road dust contamination due the vehicular traffic. The findings of present study are in line with studies conducted in cities, such as Tehran. In studies carried out in Tehran, the ERI was very high in all sections ³¹.

Shapiro-Wilk and Kolmogorov-Smirnov tests indicated that distribution of metals concentrations was not normal. Thus, spearman correlation analysis was applied for determination of the association between heavy metals and their origin. The results showed the highest correlation coefficients ($r > 0.6$) for metals, including Cu, Fe, Ni, Sb, Zn, and U, indicating that they originate from the same anthropogenic source.

PCA establishes further data about the possible correlation between heavy metals and their sources ⁴¹. The classification of Ni, Cr, Fe, and Mn suggests geogenic sources. The second PC (PC2) representing 22.00% of the total variance shows high level of As, Sb, U, Pb, and Mo (metals highly related to anthropogenic sources).

Positive matrix factorization (PMF) model was performed to investigate the contribution of various sources of heavy metals in road dust of Behbahan city. Previous studies reported that Cu, Zn, and Pb in road dust might be due to traffic emission. The important sources of Cu, Pb, and Zn in road dusts were petrol combustion by vehicles. Additionally, high level of Pb and Zn in road dust might be related to fossil combustion and industrial practice ⁴². All of these metals (Pb, Cu, and Zn) are related to road pavement materials, tire and brake wear, lubricants, and fuel combustion. Previous studies indicated that Pb, Cr, As, Mo, and Sb originated from industrial area ⁴³. Another study investigated that source of Cd and Cr was from

industrialization and human activities ⁴⁴. Anthropogenic source (third factor) was prevailed by Fe (52%), Al (49%), Ni (56.2 %), Mn (58.9 %), Co (44.3 %), and Cr (43.9%). Al and Fe indicated that common geochemical commitment of significant stone framing components is in industrial region soils.

Because of children's physiological and behavioral characteristics, such as high respiration rates, hand to mouth activities, and high gastrointestinal absorption, they have a higher sensitivity of exposure to pollutants than adults ⁴⁵. The average carcinogenic risks of carcinogen metals (As, Co, Cr, Ni, and Pb) under different routes for adults and children were below 10^{-2} . Furthermore, TCR for adults and children were 6.89×10^{-3} and 4.02×10^{-2} , respectively, demonstrating that human exposure to road dust in the study area may not pose significant health effects to adults, but may pose carcinogenic effects to children.

Conclusion

In this study, the concentration of heavy metals in road dust of Behbahan city was investigated. The concentration of heavy metals, such as Cu, Zn, Ni, As, Cr, Pb, U, and Fe in the road dust of the commercial zone were much higher than other zones. The concentrations of heavy metals, such as Co and Sb in the heavy traffic section were much higher than other sections. These metals vary widely between different sections and between samples, due to heterogeneity of samples and various variables that cannot be controlled. The highest I_{geo} values for Pb, Cu, and Sb were 1.39, 2.19, and 2.46, which may suggest that Pb, Cu, and Sb in the road dust were most affected by anthropogenic sources. The lowest ecological risk was determined in residential zone. Heavy traffic zone shows the maximum RI value and ecological risk of the road dust samples. The results indicated that chronic exposure of humans (both adults and children) to road dust in the study area may pose adverse health effects. Furthermore, TCR for adults and children were 6.89×10^{-3} and 4.02×10^{-2} , respectively, indicating that human exposure to

road dust in the study area may not pose significant health effects to adults, but may pose carcinogenic effects to children. The results reported the highest correlation coefficients ($r > 0.6$) for metals, such as Cu, Fe, Ni, Sb, Zn, and U that represent they originated from the same anthropogenic source.

Acknowledgements

Thanks are owed to National Institute for Medical Research Development (NIMAD) for financially supporting this study with the grant number 971551.

Funding

This study was funded by National Institute for Medical Research Development (NIMAD).

Conflict of interests

The authors declare no conflict of interest.

This is an Open-Access article distributed in accordance with the terms of the Creative Commons Attribution (CC BY 4.0) license, which permits others to distribute, remix, adapt, and build upon this work for commercial use.

References

1. Akhter MS, Madany IM. Heavy metals in street and house dust in Bahrain. *Water Air Soil Pollut.* 1993;66(1-2):111-9.
2. Lu X, Wang L, Lei K, et al. Contamination assessment of copper, lead, zinc, manganese and nickel in street dust of Baoji, NW China. *J Hazard Mater.* 2009;161(2-3):1058-62.
3. Roger S, Montrejaud-Vignoles M, Andral M, et al. Mineral, physical and chemical analysis of the solid matter carried by motorway runoff water. *Water Res.* 1998;32(4):1119-25.
4. Núñez O, Fernández-Navarro P, Martín-Méndez I, et al. Association between heavy metal and metalloid levels in topsoil and cancer mortality in Spain. *Environ Sci Pollut Res.* 2017;24(8):7413-21.
5. Acosta PA, Lartey EK, Mandelman FS. Remittances and the Dutch disease. *J Int Econ.* 2009;79(1):102-16.
6. Zheng N, Liu J, Wang Q, et al. Health risk assessment of heavy metal exposure to street dust in the zinc smelting district, Northeast of China. *Sci Total Environ.* 2010;408(4):726-33.
7. Banerjee AD. Heavy metal levels and solid phase speciation in street dusts of Delhi, India. *Environ Pollut.* 2003;123(1):95-105.
8. Harrison RM, Laxen DP, Wilson SJ. Chemical associations of lead, cadmium, copper, and zinc in street dusts and roadside soils. *Environ Sci Technol.* 1981;15(11):1378-83.
9. Wuana RA, Okieimen FE. Heavy metals in contaminated soils: a review of sources, chemistry, risks and best available strategies for remediation. *ISRN Ecology.* 2011;2011:1-21.
10. Kim J, Park J, Hwang W. Heavy metal distribution in street dust from traditional markets and the human health implications. *Int J Environ Res.* 2016;13(8):820.
11. Duong TT, Lee BK. Determining contamination level of heavy metals in road dust from busy traffic areas with different characteristics. *J Environ Manage.* 2011;92(3):554-62.
12. Kim KW, Myung JH, Ahn J, et al. Heavy metal contamination in dusts and stream sediments in the Taejon area, Korea. *J Geochem Explor.* 1998;64(1-3):409-19.
13. Joshi A, Roh H. The role of context in work team diversity research: A meta-analytic review. *Acad Manag Ann.* 2009;52(3):599-627.
14. Hakanson L. An ecological risk index for aquatic pollution control. A sedimentological approach. *Water Res.* 1980;14(8):975-1001.
15. Tomlinson D, Wilson J, Harris C, et al. Problems in the assessment of heavy-metal levels in estuaries and the formation of a pollution index. *Helgol Mar Res.* 1980;33(1):566.
16. Muller G. Index of geoaccumulation in sediments of the Rhine River. *Geo Journal.* 1969;2:108-18.
17. USEPA. Risk Assessment Guidance for Superfund-Volume I: Human Health Evaluation Manual (Part E, Supplemental Guidance for Dermal Risk Assessment) Interim Review Draft. Office of Emergency and Remedial Response Washington, DC, USA; 2001.

18. Ferreira-Baptista L, De Miguel E. Geochemistry and risk assessment of street dust in Luanda, Angola: a tropical urban environment. *Atmos Environ*. 2005;39(25):4501-12.
19. Kong S, Lu B, Bai Z, et al. Potential threat of heavy metals in re-suspended dusts on building surfaces in oilfield city. *Atmos Environ*. 2011;45(25):4192-204.
20. Amato F, Pandolfi M, Moreno T, et al. Sources and variability of inhalable road dust particles in three European cities. *Atmos Environ*. 2011;45(37):6777-87.
21. Ram S, Kumar R, Chaudhuri P, et al. Physico-chemical characterization of street dust and re-suspended dust on plant canopies: an approach for finger printing the urban environment. *Ecol Indic*. 2014;36:334-8.
22. Lu S, Bai S, Xue Q. Magnetic properties as indicators of heavy metals pollution in urban topsoils: a case study from the city of Luoyang, China. *Geophys J Int*. 2007;171(2):568-80.
23. Trujillo-González JM, Torres-Mora MA, Keesstra S, et al. Heavy metal accumulation related to population density in road dust samples taken from urban sites under different land uses. *Sci Total Environ*. 2016;553:636-42.
24. Soltani N, Keshavarzi B, Moore F, et al. Ecological and human health hazards of heavy metals and polycyclic aromatic hydrocarbons (PAHs) in road dust of Isfahan metropolis, Iran. *Sci Total Environ*. 2015;505:712-23.
25. Iqbal J, Shah MH. Study of selected metals distribution, source apportionment, and risk assessment in suburban soil, Pakistan. *J Chem*. 2015;2015:1-8.
26. Ravankhah N, Mirzaei R, Masoum S. Evaluation of geoaccumulation index, contamination factor, and principal component analysis for estimating soil contamination. *Iranian Journal of Health, Safety and Environment*. 2015;8(3):345-56.
27. Sun Y, Zhou Q, Xie X, et al. Spatial, sources and risk assessment of heavy metal contamination of urban soils in typical regions of Shenyang, China. *J Hazard Mater*. 2010;174(1-3):455-62.
28. Yongming H, Peixuan D, Junji C, et al. Multivariate analysis of heavy metal contamination in urban dusts of Xi'an, Central China. *Sci Total Environ*. 2006;355(1-3):176-86.
29. Wei B, Jiang F, Li X, et al. Spatial distribution and contamination assessment of heavy metals in urban road dusts from Urumqi, NW China. *Microchem J*. 2009;93(2):147-52.
30. Martínez LLG, Poletto C. Assessment of diffuse pollution associated with metals in urban sediments using the geoaccumulation index (I_{geo}). *J Soils Sediments*. 2014;14(7):1251-7.
31. Saeedi M, Li LY, Salmanzadeh M. Heavy metals and polycyclic aromatic hydrocarbons: pollution and ecological risk assessment in street dust of Tehran. *J Hazard Mater*. 2012;227:9-17.
32. Apeagyei E, Bank MS, Spengler JD. Distribution of heavy metals in road dust along an urban-rural gradient in Massachusetts. *Atmos Environ*. 2011;45(13):2310-23.
33. Fazeli MS, Moosavi M, Pournia M, et al. Metals distribution in topsoils around industrial town of Ahwaz II, Ahwaz, Iran. *Res J Appl Sci*. 2009;9(6):1121-7.
34. Song D, Zhuang D, Jiang D, et al. Integrated health risk assessment of heavy metals in Suxian County, South China. *Int J Environ Res*. 2015;12(7):7100-17.
35. Malekpouri M, Ehsanpour M, Afkhami M. Metal levels in airborne particulate matter in industrial area of bandar abbas, Iran. *Eur J Exp Biol*. 2012;2(5):1714-7.
36. Li X, Liu L, Wang Y, et al. Heavy metal contamination of urban soil in an old industrial city (Shenyang) in Northeast China. *Geoderma*. 2013;192:50-8.
37. Okorie A, Entwistle J, Dean JR. Estimation of daily intake of potentially toxic elements from urban street dust and the role of oral bioaccessibility testing. *Chemosphere*. 2012;86(5):460-7.
38. Wei B, Jiang F, Li X, et al. Heavy metal induced ecological risk in the city of Urumqi, NW China. *Environ Monit Assess*. 2010;160(1-4):33.

39. Wei B, Yang L. A review of heavy metal contaminations in urban soils, urban road dusts and agricultural soils from China. *Microchem J*. 2010;94(2):99-107.
40. Rahmani H, Kalbasi M, Hajrasuliha S. Lead-polluted soil along some Iranian highways. *Journal of Crop production and processing*. 2001;4(4):31-42.
41. Akhbarizadeh R, Moore F, Keshavarzi B, et al. Microplastics and potentially toxic elements in coastal sediments of Iran's main oil terminal (Khark Island). *Environ Pollut*. 2017;220:720-31.
42. Bhuiyan MAH, Dampare SB, Islam M, et al. Source apportionment and pollution evaluation of heavy metals in water and sediments of Buriganga River, Bangladesh, using multivariate analysis and pollution evaluation indices. *Environ Monit Assess*. 2015;187(1):1-21.
43. Jiang Y, Chao S, Liu J, et al. Source apportionment and health risk assessment of heavy metals in soil for a township in Jiangsu Province, China. *Chemosphere*. 2017;168:1658-68.
44. Luo L, Ma Y, Zhang S, et al. An inventory of trace element inputs to agricultural soils in China. *J Environ Manage*. 2009;90(8):2524-30.
45. Chen H, Teng Y, Lu S, et al. Source apportionment and health risk assessment of trace metals in surface soils of Beijing metropolitan, China. *Chemosphere*. 2016;144:1002-11.